

1 Effects of Preconditioning the Rhizosphere of
2 Different Plant Species on Biotic Methane
3 Oxidation Kinetics

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Ndanga, É., Lopera, C., Bradley, R. L. and Cabral, A.R. (2016). Effects of Preconditioning the Rhizosphere of Different Plant Species on Biotic Methane Oxidation Kinetics. *Waste Management*, 55: 313-320. DOI: 10.1016/j.wasman.2016.04.035.

9 **ABSTRACT**

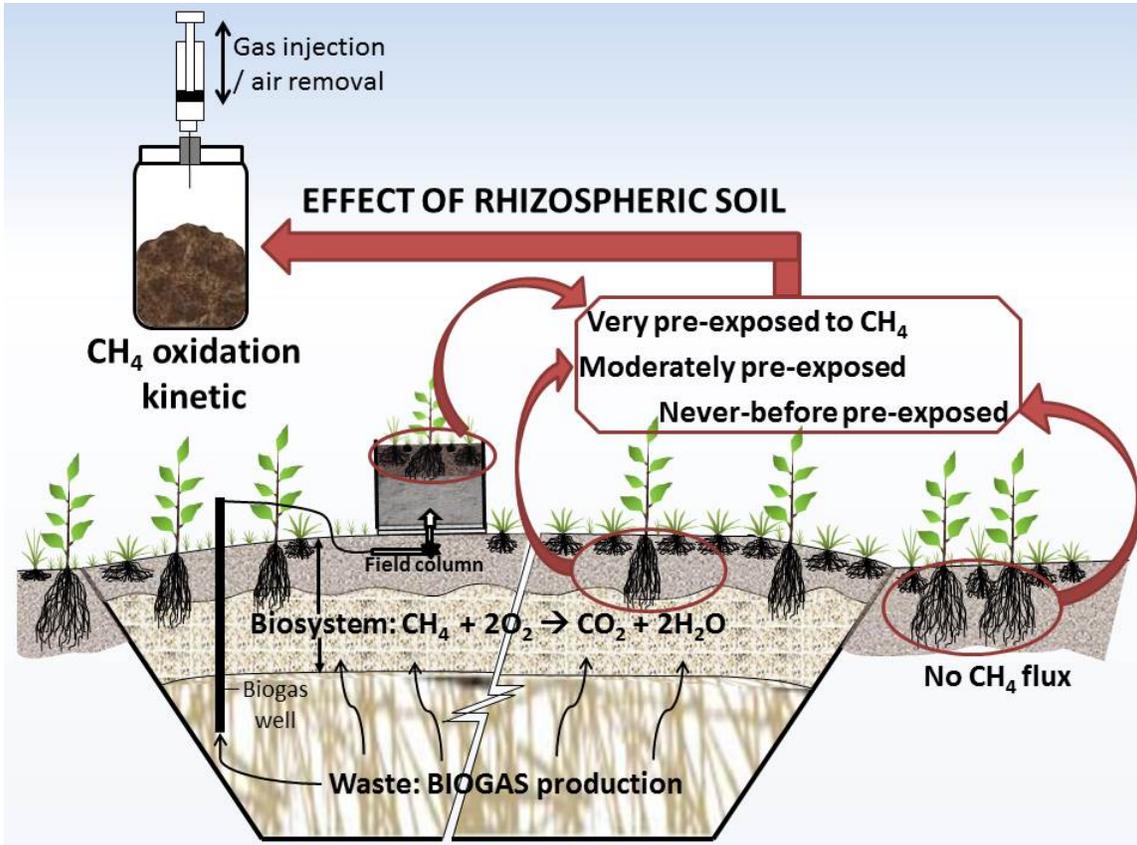
10 The rhizosphere is known as the most active biogeochemical layer of the soil. Therefore,
11 it could be a beneficial environment for biotic methane oxidation. The aim of this study
12 was to document - by means of batch incubation tests - the kinetics of methane oxidation
13 in rhizospheric soils that were previously exposed to methane. Three CH₄ pre-exposure
14 zones were sampled; the never-before pre-exposed, the moderately pre-exposed and the
15 very pre-exposed. For each zone, 3 different plant species were considered. All of the
16 samples were placed in Mason jars and submitted to the same initial CH₄ concentration.
17 CH₄ consumption started in less than 3 days and ended in less than a week. The results
18 showed that the fastest CH₄ consumption occurred for the rhizospheric soil that was
19 moderately pre-exposed. However, no statistically significant differences were found in
20 the CH₄ oxidation kinetic parameters (lag time and half-life) of all the rhizospheric soils,
21 suggesting that methane oxidation did not depend on plant species or CH₄ pre-exposure
22 levels, for the soils and plants tested herein. The oxidation rate values obtained were
23 higher than those reported in the reviewed literature for unplanted landfill cover soils.

24 **KEYWORDS:** Rhizosphere, methane oxidation, kinetics, vegetation, preconditioning

25

26 TOC/ABSTRACT ART.

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30 INTRODUCTION

31 Biotic methane oxidation is a highly efficient biogeochemical process to reduce CH₄
32 emissions from landfills through the use of engineered biosystems. These biosystems are
33 typically made up of a sequence of soil layers that promote the development of methane
34 oxidizing bacteria, which use CH₄ as their energy source.¹

35 Although methane oxidation (CH₄-ox) has been extensively documented, there are a
36 limited number of studies on the effects of vegetation on methanotrophic activity and
37 CH₄ oxidation efficiency in landfills. Among these studies, Bohn et al.² and Reichenauer
38 et al.³ assessed the impact of several types of plant covers on CH₄ oxidation. Their
39 studies concluded that CH₄ oxidation efficiencies of the tested biosystems differed, and
40 that vegetation enhanced biotic CH₄ oxidation. Wang et al.⁴ evaluated how *Chenopodium*
41 *album* L. affected methanotrophic activity. They observed a significant increase in the
42 total number of soil culturable bacteria in soils seeded with *C. album* L. and exposed to
43 landfill gas. According to Wang et al.⁴, the total number of methanotrophic bacteria in the
44 seeded soils exposed to landfill gas was significantly higher than in soils either not
45 exposed to landfill gas or seeded. In a recent study, Ndanga et al.⁵ reported the impact of
46 *Trifolium repens* L., *Phleum pratense* L., and a mixture of both on aerobic CH₄ oxidation.
47 Ndanga et al.⁵ performed column studies in the laboratory and in the field and concluded
48 that the influence of vegetation on methane oxidation was not noticeable for loadings up
49 to approximately 100 g m⁻² d⁻¹. In fact, these authors observed that the bare soil
50 performed as well as the 3 vegetated covers they analyzed. In the above mentioned

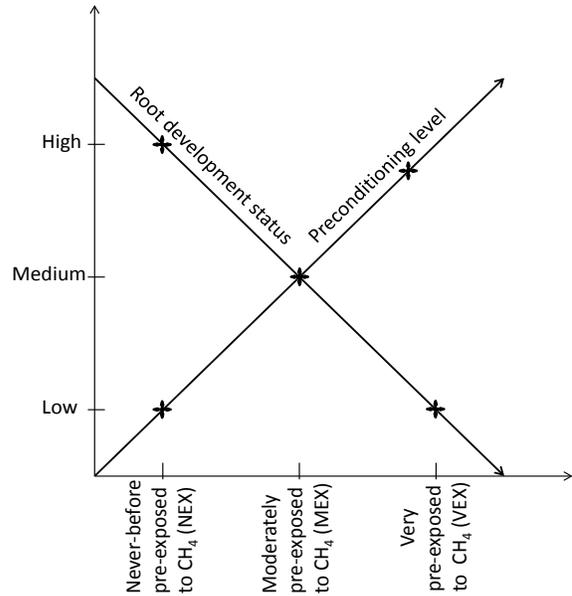
51 studies, it was concluded that a positive impact of vegetation on biotic CH₄ oxidation
52 could be attributed to enhanced nutrient supply through the root system and gas diffusion.

53 The effect of vegetation on CH₄ oxidation in landfills may also be assessed through the
54 efficiency of the rhizospheric soil in oxidizing CH₄. This is another subject that remains
55 poorly documented in the technical literature relating to biosystems incorporated in
56 landfill final covers. The rhizosphere is the zone of contact between root and soil that
57 supports high levels of bacterial activity. Through root exudations (rhizodeposition),
58 plants release organic compounds such as amino acids and sugars, which serve as energy
59 substrates for soil heterotrophic microorganisms. This, in turn, may stimulate soil nutrient
60 turnover with positive effects on methanotrophic populations. Accordingly, the
61 abundance of micro-organisms in the rhizosphere is generally 5 to 20 times greater than
62 that found in non-vegetated soil.^{6,7} Therefore a well-developed rhizospheric soil would
63 accelerate CH₄ oxidation. The variability in microbial composition and activity in the
64 rhizosphere generally depends on the quality and quantity of root exudations, which vary
65 according to plant species, age and vigor, as well as site specific factors such as soil
66 chemical properties and climatic conditions.⁸

67 The aim of the present study was to assess the importance of the rhizosphere in methane
68 oxidation kinetics for a limited number of plant species commonly found at the Saint-
69 Nicephore landfill in Quebec, Canada. We hypothesized that preconditioning the
70 rhizospheric soils (therefore the microflora) through exposure to CH₄ (in fact landfill gas)
71 might lead to faster development of CH₄ oxidizing bacteria and, as a consequence, to
72 faster and earlier CH₄ oxidation, in comparison with unconditioned rhizospheric soils.

73 This experimental plan was designed to document CH₄ oxidation within the rhizosphere
74 layer of the biosystems tested. The rhizospheric soils associated with 3 levels of root
75 development, and for different plant species, were exposed to 3 levels of CH₄
76 preconditioning. As schematized in Figure 1, the lowest, medium and highest levels of
77 root development were found in the very pre-exposed, moderately pre-exposed and
78 never-before pre-exposed soils respectively. The kinetic approach, which has been
79 commonly employed to assess the capacity of soil materials to oxidize CH₄, was adopted
80 herein. It involves batch incubation for determining the kinetic parameters, such as the
81 potential rate of CH₄ oxidation and its associated half-life ($t_{1/2}$). To determine the CH₄-ox
82 kinetic parameters, rhizospheric soil samples were collected from 2 selected locations
83 within a covered landfill and 1 from its surroundings, each associated with 1 of the 3
84 levels of preconditioning.

85 One practical outcome of this investigation was the possibility of assessing whether or
86 not some level of exposure of the rhizospheric soil to methane, prior to construction of
87 the biosystem, might result in a faster - and more substantial - reduction in fugitive
88 methane fluxes from landfills.



89

90 **Figure 1.** Schematic framework of the study

91

92 **MATERIALS AND METHODS**

93 **Study site**

94 Samples were obtained at a landfill site located in the municipality of Drummondville
 95 (Saint-Nicéphore), Quebec, Canada. Sampling regions were chosen based on plant cover
 96 and CH₄ pre-exposure level. 3 CH₄ pre-exposure zones were selected: 1) never-before
 97 pre-exposed zone (NEX), located < 30 m away from the landfill site, so as to obtain the
 98 same plant species found on the final cover; 2) moderately pre-exposed zone (MEX),
 99 located in an area where the final cover was placed approximately 8 years ago, and for
 100 which CH₄ emissions oscillate between 0-200 ppm; 3) very pre-exposed zone (VEX).
 101 The latter was not actually a zone, but our own field column experiments, described in
 102 Ndanga et al.⁵

103 For each CH₄ pre-exposure zone, 3 plant species were considered for the study: clover
104 (*Trifolium spp.* L.), timothy grass (*Phleum pratense* L.) and perennial ryegrass (*Lolium*
105 *perenne* L.). These are commonly found species in the area within and around the
106 landfill. However, for the VEX-soil, instead of ryegrass, a mixture of clover and timothy
107 grass was used. As shown in Figure 2, all 9 treatments (3 CH₄ pre-exposure levels x 3
108 plant species) were carried out in 5 replicates, or blocks, for a total of 45 samples.

109 For all the CH₄ pre-exposure levels, sampling points were defined based on the premise
110 that, in a delimited area of 1.5 x 1.5 m, a sampling point was retained if the plant species
111 of interest for the study covered at least 75% of total plant cover in the area. The
112 sampling points chosen were geo-referenced using a GPS.

113 **Soil sampling**

114 Rhizospheric soil is defined as the soil that adheres to the roots after gently shaking them.
115 Rhizospheric soil samples were collected in the upper 15-cm layer in July 2013, placed in
116 sterilized plastic bags and kept at approximately 4°C until further processing. The soils
117 were classified as fine sand and their natural physical characteristics (water content and
118 organic matter) were measured. CH₄ incubation tests were performed without any
119 adjustment of these initial values (Table 1).

120

121 **Table 1.** Physical characteristics of rhizospheric soil

Soil type	Plant species	Water content (w/dw %) ¹	Organic matter (%) ¹
NEX	Clover	11.3 ± 3.2	3.9 ± 0.7
	Timothy grass	13.9 ± 1.9	4.4 ± 0.7
	Ryegrass	10.0 ± 3.4	3.2 ± 1.5
MEX	Clover	20.9 ± 2.5	4.9 ± 1.4
	Timothy grass	21.2 ± 5.3	4.6 ± 1.4
	Ryegrass	26.2 ± 8.6	5.6 ± 0.9
VEX	Clover	21.3 ± 13.4	10.6 ± 3.0
	Timothy grass	27.8 ± 7.8	10.3 ± 1.7
	Mixture	22.7 ± 9.2	8.9 ± 1.8

122 ¹ Data are the mean ± standard deviation of the 5 replicates of the rhizospheric soil.

123 **CH₄-ox kinetic measurements**

124 After rhizospheric soil samples were removed from the refrigerator, they were left
 125 overnight at ambient temperature before incubation. Approximately 100 g of fresh soil
 126 samples were incubated in 500-ml Mason jars tightly closed. The jars were previously
 127 tested to confirm their air-tightness (no loss of gas). A rubber septum was fitted to the cap
 128 of the Mason jars to allow gas injection and sampling using a syringe.

129 Methane was injected into the jars so as to reach an initial CH₄ concentration of ~ 45%.

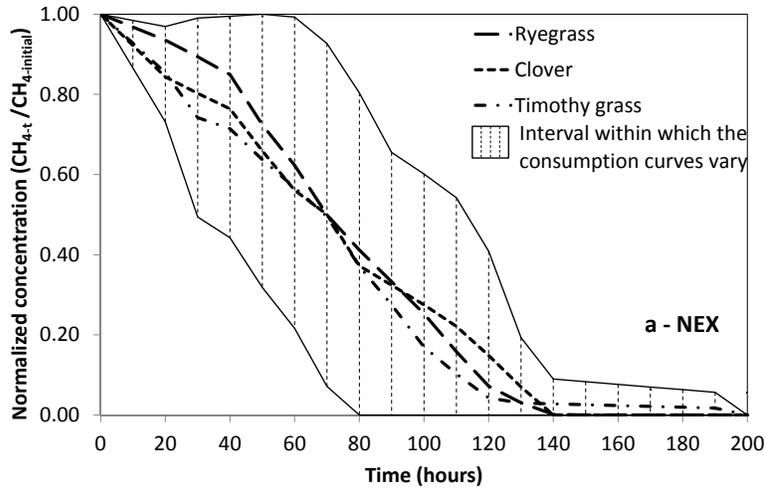
130 Before CH₄ injection, an equal amount of air was removed from the jar to equilibrate the
 131 pressure. The first gas measurement was taken within the first 2 minutes after gas
 132 injection. CH₄, CO₂ and O₂ concentrations in the headspace were monitored using a
 133 Micro GC Agilent Technologies 3000A gas chromatograph. Incubation in each jar
 134 continued until the CH₄ concentration decreased below 1%.

135 Analyses were performed one block (Figure 2) at a time and 3 parameters were defined in
136 order to describe the CH₄-ox kinetics: the lag time, which is the time required for the
137 rapid acceleration in CH₄ consumption to occur; the oxidation rate, k, and the CH₄
138 oxidation half-life (related to the oxidation rate). The latter two are respectively the
139 average slope of CH₄ consumption versus time plot, and the time required for the CH₄
140 concentration to decrease by one-half in the headspace.

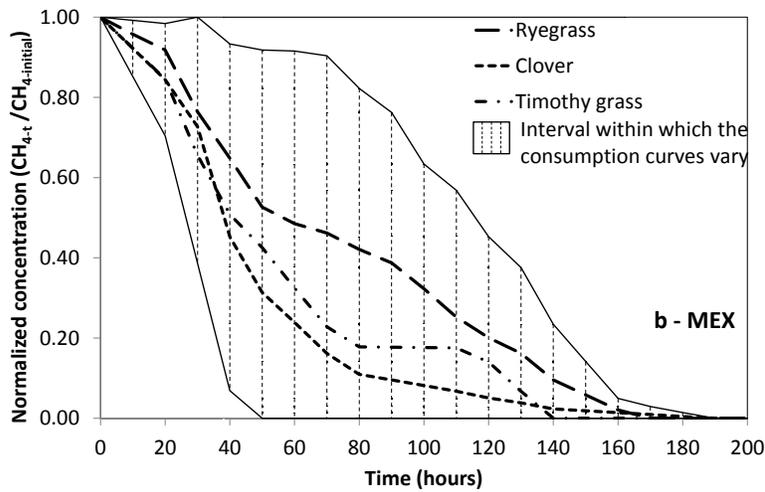
141 The oxidation rate was obtained by fitting the measured CH₄ concentrations to the typical
142 zero-order reaction equation, i.e. $-k = dC/dt$, where C is the concentration. k was
143 normalized to the soil dry weight and its unit was $\mu\text{mol CH}_4 / \text{g d.w. / h}$.

144 **Statistical analyses**

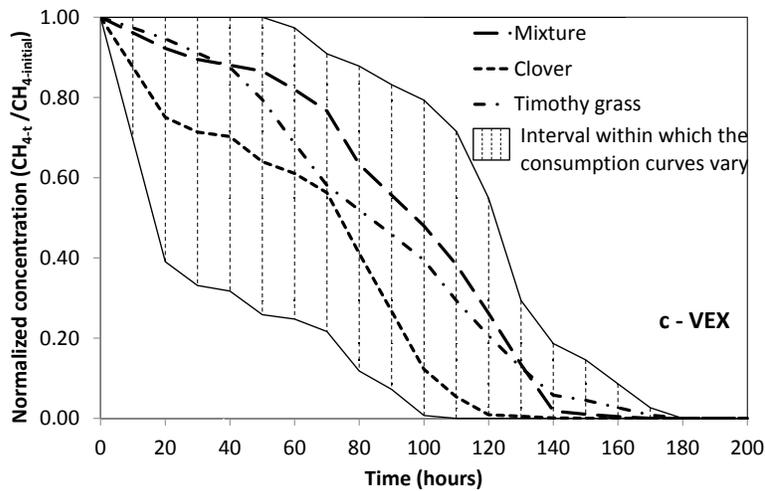
145 The results were analysed in two different ways. The first was an analysis of the effect of
146 the 9 different treatments (each representing a combination of CH₄ pre-exposure level
147 and plant species; Figure 2) on lag time and CH₄ oxidation half-life responses. It was
148 impossible to dissociate the effect of all the plant species studied herein from the effect of
149 CH₄ pre-exposure, due to the fact that the plant species were not always the same for
150 each level of CH₄ pre-exposure. A one-way ANOVA test was performed for each
151 response variable of the 9 treatments. The second analysis (two-way ANOVA) tested the
152 effect of CH₄ pre-exposure level and plant species on CH₄ oxidation half-life. Only
153 timothy grass and clover, the 2 plant species found in all 3 zones of CH₄ pre-exposure,
154 were considered in the second analysis. Post hoc Scheffé tests were performed after each
155 ANOVA to identify differences between the treatments. For each statistical analysis, the
156 statistical significance threshold was 95%.



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170

171 **Figure 3.** Variation of average normalized CH₄ concentrations with time within the
 172 rhizospheres of (a) never-before pre-exposed; (b) moderately pre-exposed; and (c) very
 173 pre-exposed soils.

174 For all the plant species studied herein, it took 50 to 200 hours for the rhizospheric soil to
175 oxidize all the CH₄. It is noteworthy that the average normalized methane concentration
176 curves started to level off approximately at the same time in all cases. The average time
177 for total CH₄ consumption was 140 hours. Regardless of plant species, the average curves
178 for the NEX soils nearly superimposed, suggesting that the plant species studied herein
179 did not affect the CH₄ response of NEX soils, as far as CH₄ oxidation is concerned.

180 However, for the MEX and VEX soils, the curves obtained for the clover rhizosphere soil
181 were below the others, suggesting that the CH₄ consumption in this rhizosphere was
182 faster compared to timothy grass, ryegrass or mixture rhizospheres.

183 As expected, the normalized CH₄ concentration curves varied widely between replicates
184 (hatched zone). This variation was attributed to the fact that the samples were not
185 homogeneous. Variations could therefore be a consequence of the specific characteristics
186 of the samples, such as water content and organic matter content. For timothy grass, for
187 example, it took 50 hours for one replicate to consume all the CH₄, whereas for another
188 replicate, it took 190 hours.

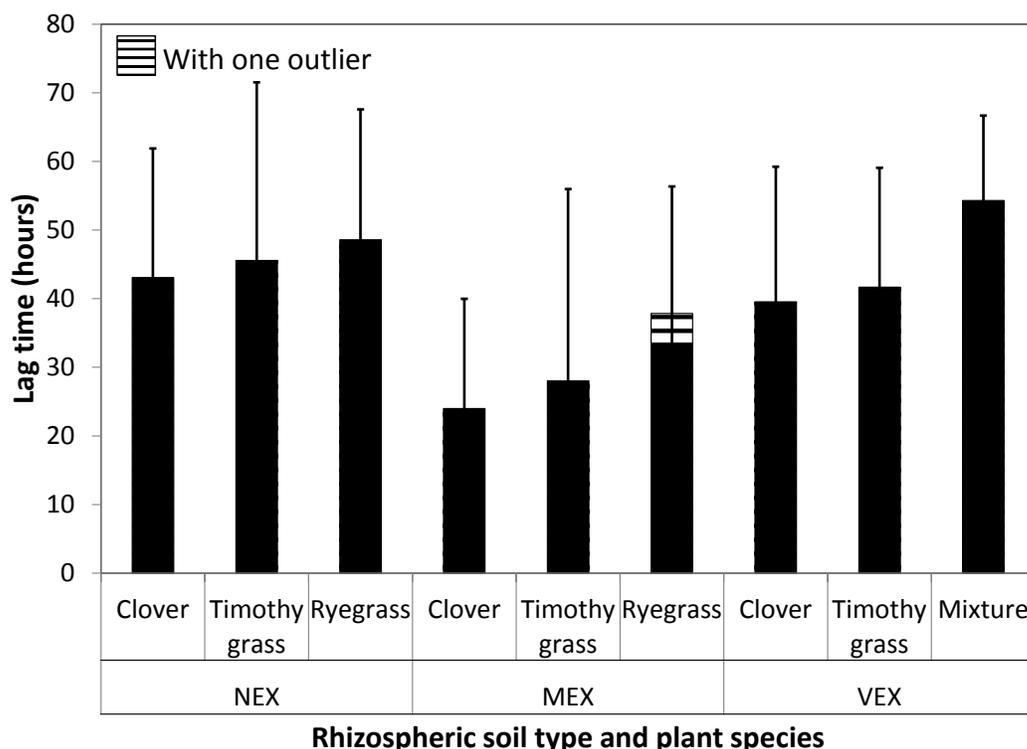
189 Figure 3 reveals 3 phases of the consumption curves: lag time, high CH₄ consumption
190 and residual consumption. The lag time, whose values are presented in Figure 4, is the
191 time required for the soil to accelerate CH₄ consumption. The high CH₄ consumption
192 phase corresponds to a period of accelerated CH₄ consumption, i.e. where the oxidation
193 rate (*k*) is the highest. The residual consumption phase is associated with the slowdown
194 in the CH₄ consumption curve. The lag time for MEX soils (29.9±22.2 h) was shorter
195 than those for the NEX and VEX soils (45.7±21.3 h and 45.2±16.5 h, respectively).

196 Moreover, Figure 3 shows that the CH₄ consumption was faster for MEX than for the two

197 other soils. Indeed, after 60 hours, 50 to 75% of CH₄ was consumed in MEX soils, while
198 only 20 to 40% was consumed in NEX and VEX soils. NEX and MEX samples were
199 taken from places where the vegetation had been established for several years. Although
200 no microbial analyses were performed for this project, one would expect to find well-
201 established root zones and microbial biomass in these soil samples. The fact that MEX
202 soils had been pre-exposed to CH₄ for several years led to shorter lag times (Figure 4),
203 while lag times for the NEX samples were longer since the samples had never been pre-
204 exposed to CH₄. In the case of VEX soils, the rhizosphere was sampled from field tests 4
205 months after the bare soil had been seeded. At the time of sampling, the vegetation was
206 well grown and the rhizosphere had been exposed for more than one week to a CH₄
207 loading of approximately 200 g CH₄ m⁻²d⁻¹; a very high loading that could characterize a
208 hotspot. Since vegetation was only 4 months old for VEX samples, one cannot expect the
209 root zone to be as mature as in the NEX and MEX samples. Consequently, the VEX
210 microbial biomass was not as developed, despite the fact that the methanotrophic
211 community was well developed.

212 An ANOVA performed on lag time data revealed that there was not a significant
213 difference in lag time between the 9 treatments ($p > 0.05$). However, a significant
214 difference was found between the clover-MEX and the mixture-VEX soils. Indeed, the
215 shortest lag time was obtained for the clover-MEX soil (23.98±16.02 h) while the longest
216 was observed for the mixture-VEX soil (54.31±12.37 h). This result suggests that the
217 rhizosphere of the clover and timothy grass mixture studied herein requires more time to
218 develop a methanotrophic community and to accelerate CH₄ oxidation than all the other
219 treatments. In a study evaluating the CH₄ oxidation efficiencies (MOE) of biosystems

220 covered with different types of plants, Ndanga et al.⁵ found that the lowest MOE was
 221 obtained for biosystems covered with the mixture of plants used in the present study,
 222 rather than more uniform vegetation (such as timothy grass).



223

224 **Figure 4.** CH₄ oxidation lag times in rhizosphere of NEX, MEX and VEX soils, for
 225 different plant species. *Error bars* are standard deviations from experiments on 5
 226 replicate soil samples

227 **Effect of plant species and CH₄ pre-exposure on CH₄-ox kinetics**

228 Table 2 shows the average of the oxidation rates (k) of the 5 replicates, and the standard
 229 deviation. The constant k represents the rate of CH₄ consumption with time. It is the
 230 average slope along the CH₄ oxidation experiment (until [CH₄] ~ 0%). Therefore, the
 231 greater the value of k, the faster the reaction. In order to determine the reaction order,

232 linear regression analyses were performed using the results of consumption tests
 233 performed with the 5 replicates of each of the 9 treatments. First and zero order kinetic
 234 models were tested. The limiting condition was the initial CH₄ concentration, which was
 235 set at 45%. Most of the treatments were fitted with the zero order kinetic model (R² was
 236 statistically significant when considering p<0.05). Accordingly, oxidation rates were
 237 calculated using the zero-order reaction model. This corroborates what is usually
 238 obtained in cases of high initial CH₄ concentration.⁹⁻¹³

239 **Table 2.** Average methane oxidation rates in never-before pre-exposed, moderately pre-
 240 exposed and very pre-exposed rhizospheric soils.

Soil type	Plant species	Oxidation rate (k) ($\mu\text{mol CH}_4/\text{g d.w./h}$) ¹
NEX	Clover	0.75 ± 0.29
	Timothy grass	0.97 ± 0.35
	Ryegrass	0.80 ± 0.30
MEX	Clover	1.21 ± 0.60
	Timothy grass	1.05 ± 0.48
	Ryegrass	0.79 ± 0.54
VEX	Clover	0.94 ± 0.65
	Timothy grass	0.63 ± 0.21
	Mixture	0.53 ± 0.12

241 ¹ Data are the mean ± standard deviation ($n = 5$).

242 The oxidation rate values obtained in this study are very close to those reported in the
 243 literature for landfill cover soils. In a review paper, Scheutz et al.¹⁴ reported oxidation
 244 rates for landfill soils in the same range as those presented above, i.e. 0.5 to 1.2 μmol
 245 CH₄/g.d.w./h (8 to 19 $\mu\text{g CH}_4/\text{g.d.w./h}$).

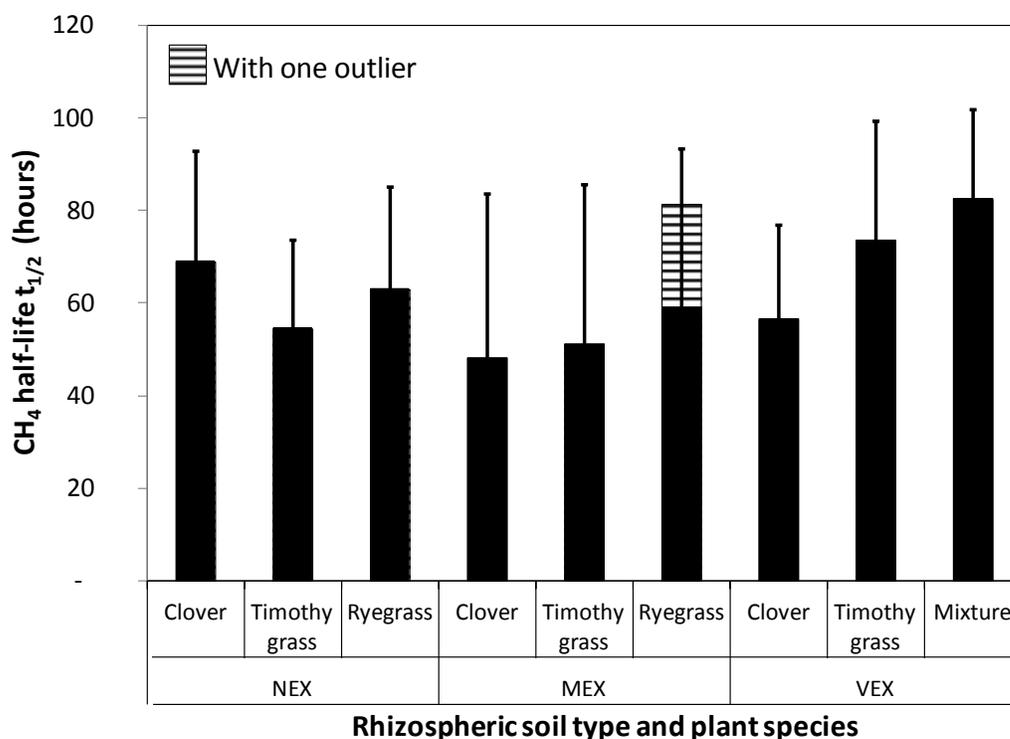
246 It can be observed in the present study that all the k values fall within the same order of
247 magnitude (Table 2). This suggests that for the soil materials studied herein, CH₄
248 oxidation was not affected by CH₄ pre-exposure level and plant species. In order to
249 confirm this observation, a one way ANOVA was performed using the methane oxidation
250 half-life, which represents the time required for the concentration of a chemical to
251 decrease by one-half. The half-life is a direct function of the oxidation rate and is
252 calculated according to the following equation:

$$t_{1/2} = \frac{0.5 \times [\text{CH}_4]_{\text{initial}}}{k} \quad [1]$$

253 Figure 5 presents the average $t_{1/2}$. The result of the ANOVA is presented in Table 3.
254 Similar to lag time, the MEX-soil had the lowest half-life values, suggesting that CH₄
255 consumption was the fastest in this rhizospheric soil. Indeed, the highest oxidation rates
256 were obtained in the MEX-soil (Table 2); specifically the clover-MEX which had the
257 shortest lag time and half-life and the highest oxidation rate. However, the statistical
258 analysis performed in all samples, including the outlier half-life value obtained for the
259 ryegrass-MEX soil, revealed that the differences between half-lives of the 9 treatments
260 were not significant ($p > 0.05$), i.e. any difference in oxidation rate cannot be attributed to
261 plant species combined to CH₄ pre-exposure, whether or not the outlier is considered. In
262 the present case differences in half-life would be attributed to factors such as initial water
263 content of the samples, organic matter content, microbial biomass and stage of root
264 development.

265 In order to distinguish the possible effect of plant species from the effect of the CH₄ pre-
266 exposure on the value of $t_{1/2}$, a two-way ANOVA was performed with clover and timothy

267 grass for the NEX, MEX and VEX soils. This analysis shows that the two factors (plant
 268 species and CH₄ pre-exposure) do not have a significant effect on t_{1/2} (p>0.05). An
 269 interaction test between these two factors on t_{1/2} also showed that the plant species did not
 270 interact (p>0.05) with CH₄ pre-exposure, as far as CH₄ oxidation half-life was concerned.



271

272 **Figure 5.** CH₄ oxidation half-lives in the rhizosphere of NEX, MEX and VEX soils for
 273 different plant species. *Error bars* are standard deviations from test results from 5
 274 replicates.

275

276 **Table 3.** Results of the ANOVA evaluating the effects of plant species and level of CH₄
 277 pre-exposure on half-life.

Source	<i>df</i>	Sum of squares	Mean Square	F statistic	p value
Block	4	6563.05	1640.76		
Model	8	6640.88	830.11	0.9426	0.496
Treatment	8	6640.88	830.11	0.9426	0.496
Residual error	32	28180.60	880.64		
Corrected total	44	41384.53			

278

279 **Further discussions and practical aspects**

280 The results presented above show that there is no clear relationship between the
 281 following parameters: 1) type of plant species considered herein (and their associated
 282 rhizospheric soil); 2) potential CH₄ oxidation in landfill covers; and 3) level of
 283 preconditioning the soil microbial communities with CH₄. In addition, Ndanga et al.⁵
 284 concluded that root depth did not influence the rate of methane oxidation in biosystems,
 285 for the same plant species studied herein. These results do not corroborate those obtained
 286 by Epp and Chanton¹⁵, Watson et al.¹⁶ and Nouchi et al.¹⁷ among others, who concluded
 287 that there is a positive correlation between methane oxidation, root depth and plant
 288 species in wetlands, peat and other anoxic media.

289 In a study comparing the CH₄-ox kinetics of a rhizospheric clayey soil (planted) and a
 290 non-rhizospheric clayey soil (unplanted), both collected from a landfill cover, Wang et
 291 al.⁴ found no significant difference in oxidation rates, except for one sampling day. They
 292 concluded that “plant growth plays an integrated role in enhancing the number and
 293 activity of soil methanotrophic bacteria”. Furthermore, several authors have reported
 294 maximum k values for unplanted sandy loam and forest soils lower than those obtained in

295 this study during the high CH₄ consumption phase (obtained from the maximum slopes of
296 the curves presented in Figure 3).¹⁸⁻²⁴ This suggests that rhizospheric soils may have
297 promoted CH₄ oxidation by providing an environment conducive for microbial growth
298 and activity. However, for mature compost Mor et al.²⁵ and Scheutz et al.¹¹ reported
299 oxidation rate values two to five times greater. These findings are not surprising given
300 that mature well-structured compost was identified and recognized as the best soil for
301 CH₄ uptake.¹⁴

302 It is noteworthy that for the high initial CH₄ concentrations applied (45%), the lag time
303 for all soils was less than 3 days. The lowest average lag times were obtained for
304 moderately pre-exposed rhizospheric soil to CH₄. The latter may have created a
305 favourable environment for methanotrophic development. There would be interest in
306 further documenting the role of pre-exposure, because it could lead to novel strategies (or
307 complementary measures) to mitigate hotspots in final covers, such as using previously
308 exposed soils around hotspots.

309 The main limitations of this study were the number of plant species tested. It would be
310 relevant to further study the interactions between vegetation and methane oxidation
311 including complex phenomena such as changes in gas diffusion patterns, changes in the
312 physical and chemical characteristics of soils due to plant growth, etc.

313

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318 **Note:** The authors declare no competing financial interest.

319 **ACKNOWLEDGEMENTS**

320 This study received financial support from the Natural Science and Engineering Research
321 Council of Canada (NSERC) and Waste Management (WM Quebec Inc.), under the
322 collaborative research and development grant # CRD 379885-08 and from the
323 Consortium de recherche et innovations en bioprocédés industriels du Québec (CRIBIQ).
324 The invaluable help of Jean-Guy Lemelin, technician, must also be acknowledged.

325 **REFERENCES**

- 326 1. Hanson, R. S.; Hanson, T. E., Methanotrophic bacteria [Review]. *Microbiological*
327 *Reviews* **1996**, *60*, (2), 439 ff.
- 328 2. Bohn, S.; Brunke, P.; Gebert, J.; Jager, J., Improving the aeration of critical fine-
329 grained landfill top cover material by vegetation to increase the microbial methane
330 oxidation efficiency. *Waste Management and Research* **2010**, *10*.
- 331 3. Reichenauer, T. G.; Watzinger, A.; Riesing, J.; Gerzabek, M. H., Impact of
332 different plants on the gas profile of a landfill cover. *Waste Management* **2011**, *31*, (5),
333 843-853.

- 334 4. Wang, Y.; Wu, W.; Ding, Y.; Liu, W.; Perera, A.; Chen, Y.; Devare, M., Methane
335 oxidation activity and bacterial community composition in a simulated landfill cover soil
336 is influenced by the growth of *Chenopodium album* L. *Soil Biol. and Bioch.* **2008**, *40*,
337 (9), 2452-2459.
- 338 5. Ndanga, É. M.; Cabral, A. R.; Bradley, R.; Johnson, T. R., Potential effect of
339 vegetation on methane oxidation efficiency of biocovers: Laboratory and field
340 experiment, In *Proceedings of the Fourteenth International Waste Management and*
341 *Landfill Symposium*, Cossu, R., Ed. CISA Publisher: S. Margherita di Pula, Cagliari,
342 Italy, 30 Sept-4 Oct, 2013; Paper 214.
- 343 6. Atlas, R. M.; Bartha, R., *Microbial Ecology: Fundamentals and Applications*. 3rd
344 ed.; The Benjamin Cummings Publishing Company, Inc: Redwood City, 1993; p 563.
- 345 7. Katznelson, H., The "rhizosphere effect" of mangels on certain groups of soil
346 microorganisms. *Soil Science* **1946**, *62*, (5), 343-354.
- 347 8. Walton, B. T.; Guthrie, E. A.; Hoylman, A. M., Toxicant degradation in the
348 rhizosphere. In *Bioremediation through Rhizosphere Technology*, American Chemical
349 Society: 1994; Vol. 563, pp 11-26.
- 350 9. Kightley, D.; Nedwell, D. B.; Cooper, M., Capacity for methane oxidation in
351 landfill cover soils measured in laboratory-scale soil microcosms. *Applied and*
352 *Environmental Microbiology* 1995, pp 592-601.
- 353 10. Bender, M.; Conrad, R., Kinetics of CH₄ oxidation in oxic soils exposed to
354 ambient air or high CH₄ mixing ratios. *FEMS Microbiology Letters* **1992**, *101*, (4), 261-
355 269.

- 356 11. Scheutz, C.; Pedicone, A.; Pedersen, G. B.; Kjeldsen, P., Evaluation of respiration
357 in compost landfill biocovers intended for methane oxidation. *Waste Management* **2011**,
358 *31*, (5), 895-902.
- 359 12. Whalen, S. C., Reeburgh, W.S. and Sandbeck, K.A., Rapid methane oxidation in
360 a landfill cover soil. *Applied and Environmental Microbiology* **1990**, *56*, 3405-3411.
- 361 13. Czepiel, P. M.; Mosher, B.; Crill, P. M.; Harriss, R. C., Quantifying the effect of
362 oxidation on landfill methane emissions. **1996**, *101*, (D11), 16721-16729.
- 363 14. Scheutz, C.; Kjeldsen, P.; Bogner, J. E.; De Visscher, A.; Gebert, J.; Hilger, H.
364 A.; Huber-Humer, M.; Spokas, K., Microbial methane oxidation processes and
365 technologies for mitigation of landfill gas emissions. *Waste Management and Research*
366 **2009**, *27*, (5), 409-455.
- 367 15. Epp, M. A.; Chanton, J. P., Rhizospheric methane oxidation determined via the
368 methyl fluoride inhibition technique. **1993**, *98*, (D10), 18413-18422.
- 369 16. Watson, A.; Stephen, K. D.; Nedwell, D. B.; Arah, J. R. M., Oxidation of methane
370 in peat: Kinetics of CH₄ and O₂ removal and the role of plant roots. **1997**, *29*, (8), 1257-
371 1267.
- 372 17. Nouchi, I.; Mariko, S.; Aoki, K., Mechanism of methane transport from the
373 rhizosphere to the atmosphere through rice plants. *Plant Physiology* **1990**, *94*, (1), 59-66.
- 374 18. Hilger, H. A.; Wollum, A. G.; Barlaz, M. A., Landfill methane oxidation response
375 to vegetation, fertilization, and liming. *Journal of Environmental Quality* **2000**, *29*, (1),
376 324-334.

- 377 19. Reay, D. S.; Nedwell, D. B.; McNamara, N.; Ineson, P., Effect of tree species on
378 methane and ammonium oxidation capacity in forest soils. *Soil Biology and Biochemistry*
379 **2005**, *37*, (4), 719-730.
- 380 20. Park, S.; Lee, C.-H.; Ryu, C.-R.; Sung, K., Biofiltration for reducing methane
381 emissions from modern sanitary landfills at the low methane generation stage. *Water, air,*
382 *and soil pollution* **2009**, *196*, (1-4), 19-27.
- 383 21. Bogner, J. E.; Spokas, K. A.; Burton, E. A., Kinetics of methane oxidation in a
384 landfill cover soil: temporal variations, a whole-landfill oxidation experiment, and
385 modeling of net CH₄ emissions. *Environ. Sci. Technol.* **1997**, *31*, (No. 9), 2504-2514.
- 386 22. Börjesson, G.; Sundh, I.; Svensson, B., Microbial oxidation of CH₄ at different
387 temperatures in landfill cover soils. *Fems Microbiology Ecology* **2004**, *48*, (3), 305-312.
- 388 23. Einola, J.-K. M.; Kettunen, R. H.; Rintala, J. A., Responses of methane oxidation
389 to temperature and water content in cover soil of a boreal landfill. *Soil Biology and*
390 *Biochemistry* **2007**, *39*, 1156-1164.
- 391 24. De Visscher, A.; Thomas, D.; Boeckx, P.; Van Cleemput, O., Methane oxidation
392 in simulated landfill cover soil environments. *Environmental Science and Technology*
393 **1999**, *33*, (1), 1854-1859.
- 394 25. Mor, S.; A., D. V.; K., R.; R.P., D.; A., C.; O., V. C., Induction of enhanced
395 methane oxidation in compost: Temperature and moisture response. *Third*
396 *Intercontinental Landfill Research Symposium* **2006**, *26*, (4), 381-388.
- 397
- 398